

Effect of Rodent Reduction on Numbers of Forest Birds in a Hawaiian Rainforest

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Introduction

Nest predation by alien rodents is one reason native bird populations in the Hawaiian Islands are decreasing (USFWS 2006). Therefore, effective mitigation of rat predation is considered an essential component of endangered species recovery programs (USFWS 2006).



Since 1996, the Maui Forest Bird Recovery Project has controlled rodents in a portion of the Hanawi Natural Area Reserve on east Maui. This program reduces black rat (Rattus rattus) numbers by 99% (Malcolm et al. 2008).

To determine if rodent reduction enhances survival and production in forest birds, we compared the

relative abundance of native and non-native species on a rodent reduction grid and in an untreated (control) area within Hanawi.

Methods

Our two study sites, HR3 and Frisbee Meadows (FSB) (Fig. 1), are located in native montane rainforest dominated by ohia (Metrosideros polymorpha). FSB has no rodent reduction. HR3 has a 26-ha grid with diphacinone bait stations and snap traps located every 50 m throughout the area. Target rodents include black rat (Rattus rattus), Polynesian rat (Rattus exulans) and Norway rat (Rattus norvegicus).

All bird abundance data were collected from January to June during 2009 and 2010 at elevations between 1500 m and 1900 m. In 2009, 101 randomly selected line transects of 500 m were completed within the HR3 rodent reduction grid. In 2010, 6 permanent transects were established. At FSB there were 6 permanent 500 m transects surveyed in both years.

Along the transects, we recorded the number of Maui Alauahio (Paroreomyza montana) and their group sizes. We expected that encounter rates and group sizes of Alauahio on the rodent reduction transects would be greater than along control transects.

To estimate relative abundance of all forest bird species, 40 standard point counts (Scott et al. 1986) were replicated in each study area, twice in 2009 and three times in 2010. Point counts result in independent birds per station (BPS) estimates, which we pooled. We completed whole model analyses with JMP software to examine count responses by year and by study site.

Common Name	Species Code	Scientific Name	Trophic Guild	Response to Rodent Reduct
Apapane	APAP	Himatione sanguinea	Nectar	None
Hawaii Amakihi	HAAM	Hemignathus virens	Invertebrates/nectar	None
liwi	IIWI	Vestaria coccinea	Nectar	Increase, P < 0.0
Akohekohe	AKOH	Palmeria dolei	Nectar	None
Maui Alauahio	MAAL	Paroreomyza montana	Invertebrates	None
Maui Parrotbill	MAPA	Psuedonestor xanthophrys	Invertebrates	None
Japanese White-eye	JAWE	Zosterops japonicus	Invertebrates	Increase, P < 0.0
Japanese Bush-warbler	JABW	Cettia diphone	Invertebrates/nectar	None
Red-billed Leiothrix	RBLE	Leiothrix lutea	Invertebrates/fruits	Increase, P < 0.0

Results

500 meter Transects



count stations. Inset of Hanawi reserve within Maui Island

In 2009, we surveyed 101 unique transects on the HR3 rodent reduction grid and repeated the 6 transects at the FSB control site at least 10 times. We converted detections to Alauahio per hour, because observers varied in the time taken to cover 500 m. Mean Alauahio per hour on the rodent reduction grid was 5.9 ± 0.3 (SE) versus 3.8 ± 0.3 (SE) in the control area (t = 4.65, p< 0.0001). Alauahio groups greater than two tended to be more prevalent on the rodent reduction grid than in the control area (Chi-square = 6.3, df = 3, p < 0.10).

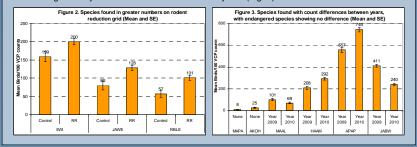
ESB

In 2010, the 6 transects at each site were sampled three times. Alauahio encounter rates showed no difference between sites and averaged 3.0 ± 0.5 (SE). Group size distributions on and off the rodent reduction grid also showed no difference in 2010.

Point counts

Nine bird species had counts sufficient for statistical analyses (Table 1).

One native and two non-native bird species responded positively to rodent reduction (RR) (Table 1, Fig. 2). Endangered species showed no difference by treatment or year (Fig. 3.) Four species differed significantly in relative abundance between years (Fig. 3).



Discussion

Transect results from 2009 are statistically superior to 2010 due to larger sample sizes. Group size differences suggest that Alauahio reproductive success is higher on the rodent reduction grid than in the control area. Additional data are desirable to reconcile the lack of replication of these results in 2010.

We selected Alauahio for monitoring on transects as a proxy for the Maui Parrotbill (Psuedonestor xanthophrvs), an endangered forest bird whose low detection rates provide little statistical inference. As expected transects and point counts did not show any difference in Parrotbill relative abundance in response to rodent reduction. However, Parrotbill pair success (number of juveniles/pair/year) improves

with rodent reduction, ranging from 0.60 to 0.80 on the HR3 grid versus 0.33 to 0.41 in the entire FSB control area (MFBRP unpublished).

According to point counts, two



non-native species benefited from rodent reduction, raising concern that interspecific competition could

affect native species. Alauahio do have diet overlap with the three non natives, and Foster (2005) found that Japanese White-eye had the most overlap. In order to have competition, food resources must be limited and reproduction must be lacking. Given Alauahio did not decline with the increase of nonnatives, competition with non-natives appears neutral. Rodent reduction should lower competition by increasing arthropod food availability (Foster 2005). It also improves forest health by reducing seed predation by rodents.

Predator control has proven to be an ally for recovering bird populations all over the Pacific (VanderWerf 2009). In order to sustain native birds for years to come, expanding rodent reduction in the Hawaiian Islands is recommended for native forest habitats. Monitoring species response to this will also need to be continued.

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